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Keywords: aqueous solution, phosphate, thermodynamic, kinetic, equilibrium, ZnO, nanoparticles

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In this study, ZnO nanoparticles were fabricated by using the hydrothermal method for adsorption of phosphate from wastewater. The obtained ZnO nanorods were characterized by powder X-ray diffraction spectroscopy (XRD), scanning electron microscopy (SEM), specific surface area (BET) and energy dispersive X-ray spectroscopy (EDS). The ZnO materials were applied for adsorption of phosphate from water using batch experiments. The effects of pH (4?10), adsorption time (30?240 min), the amount of adsorbent (0.1?0.7 g/L) and initial concentration of phosphate (147.637?466.209 mg/L) on the adsorption efficiency were investigated. The optimum condition was found at pH = 5 and at an adsorption time of 150 min. The adsorption was fitted well with the Langmuir isotherm and the maximum adsorption capacity was calculated to be 769.23 mg/g. These results show that ZnO nanomaterial would highly promising for adsorbing phosphate from water. The adsorption of phosphate on ZnO nanomaterials follows the isothermal adsorption model of Langmuir, Tempkin and Freundlich with single-layer adsorption. There is weak interaction between the adsorbent and the adsorbate. Phosphate adsorption of the ZnO nanomaterials follows Lagergren's apparent second-order kinetic model and was spontaneous and exothermic.

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Equilibrium, Kinetic and Thermodynamic Studies for Sorption of Phosphate from Aqueous Solutions Using ZnO Nanoparticles

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Abstract: In this study, ZnO nanoparticles were fabricated by using the hydrothermal method for adsorption of phosphate from wastewater. The obtained ZnO nanorods were characterized by powder X-ray diffraction spectroscopy (XRD), scanning electron microscopy (SEM), specific surface area (BET) and energy dispersive X-ray spectroscopy (EDS). The ZnO materials were applied for adsorption of phosphate from water using batch experiments. The effects of pH (4–10), adsorption time (30–240 min), the amount of adsorbent (0.1–0.7 g/L) and initial concentration of phosphate (147.637–466.209 mg/L) on the adsorption efficiency were investigated. The optimum condition was found at pH = 5 and at an adsorption time of 150 min. The adsorption was fitted well with the Langmuir isotherm and the maximum adsorption capacity was calculated to be 769.23 mg/g. These results show that ZnO nanomaterial would highly promising for adsorbing phosphate from water. The adsorption of phosphate on ZnO nanomaterials follows the isothermal adsorption model of Langmuir, Tempkin and the adsorbate. Phosphate adsorption of the ZnO nanomaterials follows Lagergren's apparent second-order kinetic model and was spontaneous and exothermic.

Keywords: nanoparticles; ZnO; equilibrium; kinetic; thermodynamic; phosphate; aqueous solution

1. Introduction

The wide direct bandgap and large exciton binding energy of zinc oxide (ZnO) have conferred the material unique electrical and photovoltaic properties and enabled its applications in the fields of fluorescence [1], photocatalytic catalysis [2], electric fire, gas sensors [3,4], electrochemical sensors [5] and solar cells [6]. ZnO can be synthesized by using various methods, including sputtering [7], sol-gel [8], co-precipitation [9], gel burning [10] and hydrothermal [11]. Depending on the synthesis routine, the morphologies of ZnO could be varied, ranging from nanorods [2], nanospheres [4], nanopores [1], nanoparticles [9] and many kinds of structures of nanorods, such as nanobelts, nanocombs and nanoforests [5]. In most synthesis routines, ZnO plays the role as an *n*-type semiconductor with electrical amphoteric properties [12]. In addition, selection of an appropriate synthesis method is also



important in ensuring operational ease and in avoiding the technical difficulties of the fabricating process. At the nano scale, ZnO finds a wide range of applications in environmental remediation,

process. At the nano scale, ZnO finds a wide range of applications in environmental remediation, bactericidal agents and in electronic devices, of which the use as adsorbent materials for removal of cation and anion pollutants features prominently due to its largely improved surface area compared to its bulk counterparts [13].

Treatment of phosphate represents an important aspect of wastewater treatment. Excessive phosphate causes overgrowth of photosynthetic cyanobacteria and algae and in turn accelerated eutrophication in water ecosystems [14]. Phosphorus could enter water bodies through effluents from households, industrial or agricultural activities, and exists in the form of $H_2PO_4^{-7}$, HPO_4^{-2} and PO_4^{-3} organic phosphate and polyphosphate [14–17]. In water, hydrolysis could occur, transforming polyphosphates and organic phosphate into other phosphate forms [18]. Therefore, industrial and agricultural discharge management, especially in terms of treatment of ammonium and phosphate, is pivotal in ensuring environmental stability and preventing water-derived health risks [19,20]. Because treatment of environmental pollution requires a large amount of adsorbents, scalability of a treatment method largely depends on its ability to easily fabricate bulk materials with a high adsorption efficiency. The hydrothermal method has been shown to be an appropriate method to synthesize ZnO due to a number of advantages, including inexpensive chemicals and ease of mass production. However, the maximum adsorption efficiency of fabricated ZnO requires optimization of all factors of adsorption, such as pH, time, the volume of the adsorbent, the initial concentration as well as careful investigation of the adsorption isothermal model, adsorption kinetic and adsorption thermal dynamic, which are relatively scarce in the literature [14–17].

In this report, we present the results of studies on isothermal adsorption, kinetics, thermodynamics of phosphate adsorption (PO_4^{3-}) removal in aqueous environments by using ZnO nanomaterials fabricated by the hydrothermal method.

2. Materials and Methods

2.1. Preparation of Materials

Zinc oxide nanomaterials were prepared by the hydrothermal method from a mixture of 25 mL $Zn(NO_3)_2 \ 0.1 \ M + NaOH \ 0.1 \ M + 20 \ mL \ C_2H_5OH$ (ratio C_2H_5OH :H2O = 1:1), pH = 11. Briefly, 0.6525 g of Zn $(NO_3)_2.4H_2O$ was weighed accurately and diluted to 25 mL of Zn $(NO_3)_2 \ 0.1 \ M$ to produce Solution A. An amount of 0.1 g of NaOH was dissolved with water in a 25 mL volumetric flask, forming a NaOH solution with a 0.1 M concentration (Solution B). Solution A was added to Solution B. Afterwards, 20 mL of C_2H_5OH (ratio C_2H_5OH : $H_2O = 1$: 1) was added to the reaction mixture above. The pH of the solution was maintained at 11. The reaction mixture was stirred on the stirrer for 20 min, and then the whole reaction mixture was heated in an autoclave at 180 °C for 24 h. The samples were taken for centrifugation, followed by washing to a pH = 7 to obtain a white, viscous gel. After obtaining the ZnO nanoparticles, the samples were heated in an air atmosphere at 350 °C for 10 h to remove the organic impurities.

2.2. Characterization

The crystalline structure and morphology of the sample were examined by using a powder X-ray diffraction spectrometer (XRD, X'Pert Pro Panalytical, Almelo, The Netherlands) equipped with Cu-K_{α} radiation ($\lambda = 1.5418$ Å) and an emission scanning electron microscope (SEM, S4800, JEOL, Tokyo, Japan), respectively. To determine the specific surface area, the Brunauer, Emmett and Teller (BET) method was used, using Tri Start 3000 equipment, Micromeritics. The distribution size of the sample was determined by the laser scattering method on HORIBA Laser Scattering Particle Size Distribution. The concentration of PO₄³⁻ before and after adsorption were determined by the UV-Vis method (Hitachi UH5300) at the University of Medicine and Pharmacy, Thai Nguyen University.

2.3. Assessment of the Prepared ZnO Nanopowder for Phosphorus Decontamination

A series of batch experiments was carried out to assess the adsorptivity of the as-synthesized ZnO nanoparticles against phosphorus. The stock solution containing potassium phosphate at a concentration of 1000 mg PO₄ per liter was first prepared to simulate wastewater. The synthesized ZnO nanoparticle was introduced into 25 mL of the diluted synthetic waste solution at a predetermined ratio and phosphorus concentration, and then mixed for a certain time period. The reaction took place in 50 mL polypropylene plastic vials under end-over-end shaking at 40 rpm. Samples were taken after specific intervals from the tube for the measurement of phosphorus concentrations using the ascorbic acid-molybdate blue method [21]. The method is based on the principle that ascorbic acid could reduce phosphorus with a blue color that is measurable by using a spectrophotometer at 885 nm. Adsorbed phosphorus would be quantified based on the initial (C_0) and final measured concentration (C). The contact time, initial phosphorus concentration, material dosage and solution pH were varied to find out its effect on the adsorption capacity of the prepared nano-ZnO.

2.4. Adsorption Capacity of the PO_A^{3-}

The adsorption experiments were investigated under a varying pH, time, weight of ZnO and initial (PO_4^{3-}) concentration. The fixed conditions included the room temperature and a shaking rate of 200 revolutions per minute (rpm).

The effect of the initial pH was measured by adding 0.025 g of ZnO to 50.0 mL of the PO_4^{3-} concentration of 150 mg/L, with the different pH values ranging from 1 to 10, followed by shaking for 150 min. The pH solution was adjusted with either by NaOH or HNO₃ solutions (0.1 M). The effect of the adsorption time was examined by adding 0.025 g ZnO to 50.0 mL PO_4^{3-} solution with an initial concentration of 150 mg/L, for a period that ranged from 30 to 240 min, at pH = 5. The effect of the ZnO mass was achieved by gradually increasing the mass of ZnO from 0.005 to 0.035 g into a 50.0 mL PO_4^{3-} solution with the initial concentration of 162.1978 mg/L, in 120 min, at pH = 5. The effect of the initial concentration of PO_4^{3-} was examined by adding 0.025 g ZnO to 50.0 mL PO_4^{3-} solution with the initial concentration of 162.1978 mg/L, in 120 min, at pH = 5. The effect of the initial concentration of PO_4^{3-} was examined by adding 0.025 g ZnO to 50.0 mL PO_4^{3-} solution with a concentration of PO_4^{3-} was examined by adding 0.025 g ZnO to 50.0 mL PO_4^{3-} solution with a concentration of PO_4^{3-} was examined by adding 0.025 g ZnO to 50.0 mL PO_4^{3-} solution with a concentration of PO_4^{3-} was examined by adding 0.025 g ZnO to 50.0 mL PO_4^{3-} solution with a concentration varying from 150 to 450 mg/L, in 150 min, at pH = 5. After the above processes, the samples were centrifuged at a rate of 4000 rpm for 10 min. The concentration of PO_4^{3-} before and after adsorption were determined by the UV-Vis method (Hitachi UH5300) at the University of Medicine and Pharmacy, Thai Nguyen University.

The adsorption capacity and adsorption efficiency were calculated according to the formula

$$q = \frac{(C_0 - C_{cb})V}{M} \tag{1}$$

$$H\% = \frac{(C_0 - C_{cb})}{C_0} \times 100\%$$
(2)

where V is the volume of the solution (L), M is the mass of the adsorbent (g), C_0 is the initial solution concentration (mg/L), Ccb is the solution concentration when the adsorption is at equilibrium (mg/L), q is the adsorption capacity at the time of equilibrium (mg/g), and H is the adsorption efficiency (%).

2.5. Equilibrium and Kinetic Modeling of the Phosphorus Decontamination Process onto ZnO

Equilibrium and kinetic modeling are essential in shedding light onto adsorption capacity. A number of different models were included: the pseudo first-order equation and second-order equation, as well as the Langmuir, Freundlich, Temkin, and Dubinin–Radushkevich equilibrium isotherm adsorption models.

3. Results and Discussion

3.1. Factors Affecting the Morphology and Surface Structure of ZnO Nanoparticles

3.1.1. SEM Image Analysis

The SEM images of the as-synthesized ZnO are shown in Figure 1. Visually, the ZnO nanoparticle structures are interwoven with some nanorods. This may be due to the developing reaction of the ZnO nanoparticles to the nanorod structures. Figure 1 shows that the ZnO materials are of high porosity and had particle sizes widely distributed due to the integration of the ZnO nanoparticles and ZnO nanorods. This distribution can be confirmed by the laser diffraction method.



Figure 1. SEM image of the ZnO nanomaterials.

3.1.2. Determination of Surface Area by the BET Method

The ZnO samples were measured by N₂ isotherm adsorption at -196 °C. The adsorption isotherm and N₂ adsorption are shown in the Figure 2. The BET graph of the ZnO samples, obtained from the software, is also shown in Figure 3. The surface areas of the ZnO samples are 17.05 m²/g.



Figure 2. Adsorption isotherm– N_2 adsorption of nano ZnO, in the as-synthesis condition at a temperature of 180 °C.



Figure 3. BET graph of the as-synthesized ZnO nanomaterials obtained at a synthesis temperature of 180 °C.

3.2. Structure and Morphology

3.2.1. XRD Analysis

The XRD patterns of the nano ZnO is shown in Figure 4, showing characteristic peaks at (100); (002); (101); (102); (110); (103); (112) and (201), which correspond to ZnO diffractions of its wurtzite structure.



Figure 4. Results of the XRD patterns of ZnO.

3.2.2. EDS Analysis

The chemical composition of the prepared ZnO samples was analyzed by the EDS method and the results are shown in Figure 5 and Table 1. From the EDS results, it is indicated that the surface composition of the ZnO is very pure, constituting about 99.99% of the Zn and O, and that only a small quantity of carbon impurities is present in the composition, possibly due to the incomplete absorption of CO_2 from the air or incomplete decomposition of the organic matter.



Figure 5. EDS pattern of the ZnO nanorods.

Table 1. Elemental analysis of the synthesized ZnO nanorods.

Element	Theory	Result
Zn	80.34	80.32
О	19.66	19.67
Total	100.00	99.99

3.2.3. Particle-Size Distribution by Laser Diffraction

The particle-size distribution in Figure 6, obtained by a laser diffraction analyzer, indicated that the majority of the ZnO particles, obtained at a synthesis temperature of 180 $^{\circ}$ C, were sized around 0.814 μ m. The range in which the ZnO particle size was measured was also low, suggesting that particles synthesized under this temperature condition was relatively stable in size. The first distribution peak is related to the ZnO nanoparticles. The second distribution peak is related to the nanorod.



Figure 6. The synthesized ZnO nanoparticle-size distribution, as per the laser diffraction analyzer.

3.3. Determination of the Isoelectric Point of ZnO

To determine the isoelectric point of ZnO, ten 0.1 M NaCl solutions were prepared with an initial pH (pH_i) ranging from 1 to 10. In each flask, 0.025 g of prepared ZnO material was added, followed by 50 mL of the NaCl solution prepared above. The mixture was allowed to stand for 48 h before being filtered and re-measured for pH. The results from Figure 7 suggests that the isoelectric point of ZnO is approximately 7.1. As a result, when the pH < pH_{pzc}, the surface of the ZnO material is positively charged; otherwise, the surface of the ZnO material is negatively charged.



Figure 7. The graph for determination of the isoelectric point of ZnO.

3.4. Adsorption of Phosphate by ZnO Nanoparticles

3.4.1. Effect of pH

Controlling the pH of the media is critical for the efficiency of the adsorption process as it plays a key role in mediating the adsorbate behaviors and adsorption capacity of the adsorbent. As illustrated in Figure 8, the adsorption efficiency was improved gradually and then eventually peaked when increasing the pH from 1 to 5, with the maximum phosphorus adsorption efficiency reaching approximately 90%. Afterwards, the efficiency dropped rapidly and at a pH of 10; the efficiency was reduced to 44.1%. Therefore, the pH value at which the phosphorus removal was maximal was recorded within the range of 5-6. According to the thermodynamic calculations, three forms of phosphate, including H₂PO₄⁻, HPO_4^{-2} and PO_4^{3-} , with a pK of 2.15, 7.20 and 12.33, respectively, existed in the aqueous solution and the ratio of the three forms is largely determined by pH of the solution. On the other hand, the point of zero charge (pH_{pzc}), which could influence the surface adsorption capacity and active centers, is lower than the pH when the pH was in the range of 2 to 12. As a result, the electrostatic repulsion and the competition between the OH⁻ ions in occupying the active sites are also the driving force for the adsorption against phosphate. Conversely, in the case of $pH < pH_{pzc}$, a positive charge of the sorbent surface might induce enhanced adsorption of the anions [19]. The reduced phosphorus adsorption at very high pH values could be explained by the competition of excessive OH⁻ and phosphate anions, such as $H_2PO_4^-$, HPO_4^{-2} and PO_4^{3-} , onto the ZnO adsorption sites [22]. Another explanation for the pH-dependent adsorption activity of ZnO is that protonation/deprotonation of the nano-ZnO surface is highly sensitive to the pH of the solution due to the amphoteric properties of the metal oxide surfaces. As a result, in the media that have a pH lower than the isoelectrical point of the ZnO, the adsorption of the H⁺ ions is promoted [23], thus leading to a positively charged ZnO surface and in turn facilitating the uptake of phosphate anions [24]. On the contrary, increasing the solution's pH further causes accumulation of the negative charges on both the nano-ZnO and the adsorbate, enhancing the electrical repulsive force between the two phases. Based on those observations, we proposed the adsorption mechanism of the phosphorus onto the ZnO particles as follows [25]:

$$= \operatorname{Zn-OH} + \operatorname{HPO}_{4}^{-2} \longleftrightarrow = \operatorname{Zn-PO}_{4}^{-2} + \operatorname{H}_{2}O$$
(3)



Figure 8. Effect of pH on the PO_4^{3-} adsorption efficiency.

It was shown that a higher solution pH is associated with degraded P adsorption. This suggests that the adsorption of phosphate anions could be hindered by a repulsive force caused by strong hydroxyl ions in the solution. Therefore, a pH = 5 was chosen for the next experiment.

3.4.2. Effect of Contact Time

Adsorption efficiencies with respect to contact duration are shown as in Figure 9. Generally, prolonging the stirring was positively correlated with the adsorption efficiency. In particular, a rapid adsorption improvement of PO_4^{3-} was observed in the first 150 min. Afterwards, the adsorption efficiency increased marginally, insignificantly during the later period.



Figure 9. Effect of time on the PO_4^{3-} adsorption efficiency.

The accelerated PO_4^{3-} uptake in the beginning period is possibly due to the abundance of adsorption holes situated on the ZnO material surface. As those holes were occupied during the process, stagnant efficiency was observed in the later stage and then the adsorption reached the equilibrium state where no adsorption improvement took place with prolonged time. Therefore, the time needed for adsorption equilibrium was selected as 150 min for subsequent experiments.

3.4.3. Effect of ZnO Dosage

The dosage of the adsorbent is responsible for both adsorption efficiency and capacity. Figure 10 presents the phosphorus removal with respect to varying ZnO dosages. Other parameters included a

contact time of 150 min, initial solute concentration of 50 ppm and solution pH of 5. Increasing the dosage from 0.1 to 0.7 g/L caused significant improvement in phosphorus adsorption, from 24.7% to 96.3%, possibly due to enhanced availability of adsorption active sites that facilitates P ion binding [26]. On the other hand, further increasing the ZnO dose higher than 0.5 g/L induced no clear P removal enhancement. This is possibly explained by saturated adsorption sites on the material surface caused by abundant active sites and the accumulation of material particles as effective surface area is lowered. Therefore, the optimal absorbent was selected as 0.5 g/L.



Figure 10. Effect of the adsorbent dose on the PO_4^{3-} adsorption efficiency.

3.4.4. Effect of Temperature

Figure 11 shows the effect of temperature on the adsorption efficiency of PO_4^{3-} . It can be observed that the removal of PO_4^{3-} gradually increased when elevating the temperature from 298 K to 323 K. This can be explained by the reduction in adsorption sites at higher temperatures. With the temperature increase from 298 to 323 K, the adsorption efficiency declined slowly from 93 to 89%. However, the amount of PO_4^{3-} adsorbed per unit volume of adsorbent will decrease as the temperature increases, suggesting that the adsorption of PO_4^{3-} by the ZnO nanomaterials is an exothermic process.



Figure 11. The effect of temperature on the efficiency of the PO_4^{3-} adsorption.

3.5. Adsorption Isotherm

3.5.1. Langmuir Isothermal Adsorption Model

Adsorption isothermal analysis plays a very important role for the purpose of designing experiments and manufacturing adsorbents. Experimental data are analyzed with Langmuir isothermal models because they are classical and simple for describing the equilibrium between the adsorbed ions on adsorbents and ions in solution at a constant temperature.

The Langmuir equation can be written as follows:

$$\frac{C_e}{q} = \frac{1}{q_{\max}K} + \frac{C_e}{q_{\max}}$$
(4)

where *q* is specific adsorbance, which is the number of mg of adsorbent per 1 g of adsorbent at equilibrium (mg/g); q_{max} : the maximum adsorption capacity (mg/g); C_e : the concentration of the adsorbent in solution at equilibrium (mg/L); and K: Langmuir's constant.

When substituting a and b with

$$a = \frac{1}{q_{\max}} \tag{5}$$

$$b = \frac{1}{K.q_{\text{max}}} \tag{6}$$

the above equation becomes y = ax + b. From empirical calculations, it is possible to calculate the K constant and the maximum adsorption capacity (q_{max}). The graph of the Langmuir model is presented as in Figure 12. Inferring from the results, the maximum adsorption capacity q_{max} and the constant K was 769.23 mg/g and 0.12 L/mg, respectively. Although the q_{max} values might not represent an indication of adsorption capacity in the long term, it could serve as a measure for a comparative study of adsorbents. As indicated in Table 2, the prepared ZnO material showed an adsorption capacity of 769.23 mg P/g, which far exceeded that of previously reported adsorbents, implying that ZnO may be a promising candidate adsorbent for treatment of phosphate from aqueous environments.



Figure 12. C_e/q and C_e correlations.

No.	Adsorbent	q _{max} (mg/g)	Ref.
1	ZnO	168.4	[27]
2	Zn-Al LDO (573 K, Curea $\frac{1}{4}$ 0.4 M)	232.0	[28]
3	Zn-Al LDH (Curea $\frac{1}{4}$ 0.4 M)	76.1	[28]
4	ZnO Nanorods	89.0	[25]
5	SnO ₂	21.5	[29]
6	WO ₃	19.0	[29]
7	Fe(III)–Cu(II) binary oxides	35.2	[30]
8	Silver nanoparticle-loaded activated carbon	4.5	[31]
9	Magnetite-enriched particles (MEP)	6.4	[32]
10	ZnO	769.23	Our work

Table 2. Maximum adsorption capacities of phosphorus onto various adsorbents.

The estimated results indicated that the Langmuir isotherm adsorption model described the adsorption process of PO_4^{3-} onto ZnO material, demonstrated by the high coefficient of determination (0.9983). Accordingly, the adsorption process is suggested to adhere to monolayer and chemical adsorption. Notably, those findings are larger than the previously published results [19–29].

In order to further justify the single layer adsorption of the PO_4^{3-} of ZnO, we evaluated the suitability through the R_L equilibrium parameter, which is calculated as

$$R_{\rm L} = \frac{1}{1 + K_{\rm L}C_{\rm O}}\tag{7}$$

where C_o: initial concentration of the substance (mg/g); and K_L: the Langmuir constant (L/mg).

The R_L parameter calculated from Figure 13 ranged from 0.123 to 0.256, which is lower than 1. Thus, it can be confirmed that the Langmuir model is suitable for the absorption of PO_4^{3-} by ZnO nanomaterials.



Figure 13. The dependence of R_L on C_o in the Langmuir model.

3.5.2. Freundlich Isothermal Adsorption Model

The Freundlich isothermal adsorption model is an empirical equation based on adsorption on heterogeneous surfaces of materials. The linear equation is usually expressed as

$$\log q_{\rm e} = \log k + \frac{1}{n} \log C_{\rm e} \tag{8}$$

where C_{cb} is the concentration at the time of equilibrium (mg/L); and q_e : the adsorption volume (mg/g). The constant *n* is the exponent in the Freundlich equation, which characterizes the energy heterogeneity of the adsorbed surface; and k: the Freundlich constant, to show the relative adsorption capacity of the adsorbent materials.

The calculation results of the reactive red adsorption process of PO_4^{3-} on ZnO by the Freundlich adsorption isothermal model are displayed in Figure 14.



Figure 14. Dependence of logq on $logC_e$ on PO_4^{3-} .

Figure 14 expresses the absorption of PO_4^{3-} on ZnO materials according to the Freundlich isothermal adsorption model. The calculated K_F was 275.145 (mg/g)(L/mg)^{1/n} and the constant of *n* was 4.907. The correlation coefficient, R², was 0.9697. The R² value of the Freundlich isotherm is lower than that of the Langmuir model. The value *n* gives an indication of the favorability of the adsorption process. When the value of *n* is less than 1, the adsorption is said to be unfavorable, but if it is greater than 1 it is favorable [33]. The value of *n* obtained in this study is greater than 1, which indicates that the adsorption process is favorable.

3.5.3. Dubinin-Radushkevich Isothermal Adsorption Model

The Dubinin–Radushkevich (D–R) isotherm is more inclusive than the Langmuir isotherm for it does not assume a homogenous surface or constant sorption potential [34]. It is used for the estimation of the characteristic porosity of the biomass as well as the mean free energy of the adsorption. The linear form of the D–R model is presented as follows:

$$\ln q = \ln q_{\max} - \beta \varepsilon^2 \tag{9}$$

where:

q: adsorption volume (mg/g);

*q*_{max}: maximum adsorption volume (mg/g);

β: constant of the adsorption energy (mol²/J²);

ε: Polanyi, described as follows:

$$\varepsilon = \operatorname{RTln}\left(1 + \frac{1}{\operatorname{Ccb}}\right) \tag{10}$$

where:

T: solution temperature (K); R: gas constant (8.314 \times 10⁻³ kJ/mol.K).

The value of the average adsorption energy, E (kJ/mol), can be calculated from D–R according to parameter β as follows:

$$E = \frac{1}{\sqrt{-2\beta}} \tag{11}$$

The value of the average adsorption energy indicates the nature of the adsorption process. When the E value is less than 8 kJ/mol, the adsorption process is physical adsorption and when the E value ranges from 8 to 16 kJ/mol, the adsorption is chemical.

The Dubinin–Radushkevich isothermal graph shown in Figure 15 reveals that phosphate adsorption does not follow the Dubinin–Radushkevich model, evidenced by the low R².



Figure 15. Dependence of lnq on ε^2 on PO₄³⁻ adsorption.

3.5.4. Tempkin Isothermal Adsorption Model

The isothermal model of Tempkin assumes that the adsorption heat of all the molecules on the surface of the material decreases linearly with the density of the coverage due to the interaction between the adsorbent and the substrate, and the adsorption is characterized by the uniform distribution of linked energy sources, up to a maximum number of associated energies. Temkin isotherms are represented by the following equation:

$$q = \frac{\mathrm{RT}}{\mathrm{b}_{\mathrm{T}}} \ln(\mathrm{K}_{\mathrm{T}}.\mathrm{C}_{\mathrm{cb}}) \tag{12}$$

The equation can be expressed in linear form as

$$q = B \ln K_{\rm T} + B \ln C_{\rm cb} \tag{13}$$

where

$$B = \frac{RT}{b_{\rm T}} \tag{14}$$

T: is the absolute temperature (K); R: gas constant (valued by 8.314×10^{-3} (kJ/mol.K); b_T: Tempkin constant (kJ/mol).

Determination of Temkin model parameters through plotting was described in Figure 16 and parameters of the models were summarized in Table 3.



Figure 16. Dependence of q on ln C_{cb} on the adsorption of PO_4^{3-} .

Table 3. Isothermal parameters: Langmuir, Freundlich, Tempkin and Dubinin–Radushkevich.

Isothermal Line Model	Parameter	s
. .	K _L (L/mg)	0.12
Langmuir	$q_{\rm max} ({\rm mg/g})$	769.23
	R ²	0.99
	K _F (mg/g)(mg/L) ^{1/n}	275.145
Freundlich	Ν	4.907
	R ²	0.97
Tempkin	K _T	6.887
	b _T (kJ/mol)	0.023
	R ²	0.9694
	$q_{\rm max} ({\rm mg/g})$	658.392
Dubinin-Radushkevich	$\beta (\mathrm{mol}^2/\mathrm{J}^2)$	-1.439
	R ²	0.6276
	E (kJ/mol)	0.5894

From Table 3, the value of the Tempkin constant b_T of 0.023 indicates the presence of a weak interaction between the adsorbents and supports the physical adsorption of phosphate by ZnO nanomaterials. The determination coefficient R² of the Langmuir, Freundlich and Tempkin models are nearly equivalent (0.99, 0.97 and 0.97) whereas the Dubinin–Radushkevich model is 0.628. Thus, the phosphate adsorption of the ZnO nanomaterials do not seem to follow the Dubinin–Radushkevich model, but the isothermal adsorption models of Langmuir, Freundlich and Tempkin. The adsorption of phosphate onto the ZnO nanomaterials is a single layer under the condition of a heterogeneous surface and there is a weak interaction between the adsorbent and adsorbate.

3.6. Study on the Adsorption Kinetics of PO_4^{3-} on ZnO Nanomaterials

The adsorption process of a solute could be conventionally described by the Lagergren rate equation, which could be presented in pseudo-first-order form as follows:

$$\frac{\mathrm{d}q_t}{\mathrm{d}t} = k_1 \left(q_e - q_t \right) \tag{15}$$

By taking the integral on both sides of the equation, with the boundary t from 0 to t and q from 0 to q_t , the following equation is obtained:

$$g(q_e - q_t) = \lg q_e - \frac{k_1}{2.303}t$$
(16)

where *q* is the amount of adsorbed PO_4^{3-} (mg/g). The subscript *t* and *q* denote the state at equilibrium and time *t* (min), respectively. k_1 is the rate constant. Plotting $\ln(q_e - q_t)$ against *t* gives the graph whose slope could be interpreted as the rate constant of the pseudo-first-order model.

The pseudo-second-order equation based on the adsorption equilibrium capacity can be expressed as

$$\frac{\mathrm{d}q_t}{\mathrm{d}t} = k_2 (q_e - q_t)^2 \tag{17}$$

By taking a similar approach, one could derive the following integrated linear form of the pseudo-second-order as follows:

$$\frac{t}{q_t} = \frac{1}{k_2 \cdot q_e^2} + \frac{1}{q_e}t$$
(18)

By linearly plotting time against $\log(q_e - q_t)$, the pseudo-first-order rate constant and sorption capacity could be calculated (Figure 17). On the other hand, the pseudo-second-order model could be plotted by considering t/q_t as a function of t (Figure 18). The model parameters are summarized as in Table 4. It was showed that the coefficient of determination of the pseudo-second-order is higher than that of the first-order model. In addition, the second-order model showed high agreement between the experimental ($q_{e(exp)}$) and calculated ($q_{e(cal)}$) capacities, suggesting that the second-order model would be more appropriate to describe the sorption process.



Figure 17. Pseudo-first-order non-linear method for adsorption of PO_4^{3-} onto ZnO.



Figure 18. Pseudo-second-order non-linear method for adsorption of PO_4^{3-} onto ZnO.

Table 4. Coefficients of the pseudo-first and second-order adsorption kinetic models and intraparticle diffusion model.

Concentration (mg/L)	$q_{\rm e.exp}$ (mg/g)	$q_{\rm e.cal}$ (mg/g)	k_1 (min $^{-1}$)	R ²
	Pseudo-first-order model			
51.765	29.984	20.289	0.0589	0.872
71.350	39.436	30.033	0.0564	0.927
103.546	53.768	42.897	0.0640	0.960
Pseudo-second-order model				
Concentration (mg/L)	$q_{\rm e.exp} \ ({\rm mg/g})$	$q_{\rm e.cal} \ ({\rm mg/g})$	k ₂ (g/mg.min)	R ²
51.765	30.073	30.959	0.015	0.999
71.350	39.436	40.650	0.010	0.999
103.546	53.590	54.645	0.010	0.999

Given that the adsorption process obeys the second-order apparent kinetics model of Lagergren, the activation energy of the absorption can be determined by the formula

$$k_2 = k_0 \exp(-E_a/RT) \tag{19}$$

where k_2 is adsorption rate constant (g/mg.min); k_0 is the initial rate constant; E_a is the activation energy (kJ/mol); R is the gas constant (R = 8.314 J/mol.K); and T is the absolute constant (K). In the above equation, k_2 can be replaced by $h = k_2 q_e^2$, reflecting the initial adsorption rate constant when q_t/t reaches zero, thus:

 $k_2 = h.exp(-E_a/RT)$ As a result,

$$E_a = RT (lnh - lnk_2).$$
⁽²⁰⁾

As previously reported [35], physisorption is due to weak van der Waal forces between the adsorbent and the adsorbate. The magnitude of the energy of physisorption is in the region of less than 20 kJ/mol. The same study also further stated that a covalent bond exists between the adsorbate and the adsorbent in chemisorption in which the substrate (adsorbent) is limited to monolayer coverage. The value of E, the mean free energy of the biosorption obtained in this study, as reflected in Table 5, was higher than 20 kJ/mol, and the Langmuir isotherm (monolayer coverage) best fitted the surface coverage in this study.

Concentration (mg/L)	h	k ₂ (g/mg.min)	E _a (kJ/mol)
170.165	26.150	2.074×10^{-4}	23.394
213.846	34.519	2.198×10^{-4}	23.938
317.802	52.322	2.238×10^{-4}	24.923

Table 5. Activation energy of PO_4^{3-} adsorption on ZnO nanomaterials.

3.7. Thermal Dynamic of Adbsorption of PO_4^{3-} onto ZnO Nanomaterials

The free energy (ΔG^0), enthalpy (ΔH^0) and entropy variation (ΔS^0) of the PO₄^{3–} adsorption process was calculated using the following equations:

$$K_D = \frac{q_e}{C_{\rm cb}} \tag{21}$$

$$\Delta G^0 = -RTlnK_D \tag{22}$$

$$\ln K_D = -\frac{\Delta G^0}{RT} = -\frac{\Delta H^0}{RT} + \frac{\Delta S^0}{R}$$
(23)

where:

KD: equilibrium constant;

q_e: adsorption volume at equilibrium (mg/g);

 C_{cb} : concentration of the substrate at equilibrium (mg/L);

R: gas constant;

T: temperature (K).

According to the calculated parameters in Table 6, the nature of the PO_4^{3-} adsorption process onto the ZnO material was spontaneous, evidenced by negativity of the change of the free energy (ΔG^0). The increase in ΔG^0 could be attributed to the elevating temperature under which the adsorption of phosphorus took place. On the other hand, the negative value of ΔH^0 suggests the reaction was endothermic. The negative value of ΔS^0 also indicated the stability of the solid–solution interface during the adsorption process [36].

Table 6. Values of the thermodynamic parameters (kJ/mol) for the adsorption of PO_4^{3-} onto ZnO.

T (K)	ΔG^0 (kJ/mol)	ΔH^0 (kJ/mol)	ΔS^0 (kJ/mol.K)
298	-8.73		
303	-8.63	-14.55	-0.02
313	-8.44		
323	-8.24		

4. Conclusions

A high adsorption efficiency of ZnO nanorods for phosphate ions was found by using the hydrothermal method. The ZnO showed a wurtzite structure with a specific surface area of $17.05 \text{ m}^2/\text{g}$ and its purity reached 99%. The as-synthesized ZnO particles were used in the adsorption of phosphate in aqueous media. The ZnO materials were applied for adsorption of phosphate from water using batch experiments. The effects of pH (4–10), adsorption time (30–240 min), the amount of adsorbent (0.1–0.7 g/L) and the initial concentration of phosphate (147.637–466.209 mg/L) on the adsorption efficiency were investigated. The optimum condition was found at pH = 5 and at an adsorption time of 150 min. The adsorption fitted well with the Langmuir isotherm and the maximum adsorption capacity was calculated to be 769.23 mg/g. These results show that ZnO nanomaterial is highly promising

for adsorbing phosphate from water. The adsorption of phosphate onto ZnO nanomaterials follows the isothermal adsorption model of Langmuir, Tempkin and Freundlich with single-layer adsorption. There is a weak interaction between the adsorbent and adsorbate.

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